

Combining human acceptance and habitat suitability in a unified socio-ecological suitability model: a case study of the wolf in Switzerland

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Summary

1. Habitat suitability models (HSMs) are commonly used in conservation practise to assess the potential of an area to be occupied and colonised. A major limitation of these models, however, is the omission of spatially explicit understanding of human acceptance towards the focal species. As wildlife is more and more subject to human-dominated landscapes, ignoring the sociological component will result in misrepresentation of the observed processes and inappropriate management.

2. We distributed 10 000 questionnaires across Switzerland and identified key socio-demographical factors correlated with human acceptance of the wolf. We then created a spatially explicit acceptance model based on geo-referenced socio-demographical, social and geographical information. Finally, we combined our acceptance model with a HSM to obtain a unified socio-ecological suitability model, which included human and ecological components.

3. We showed that the key factors associated with human acceptance were perception of how harmful the wolf is, interest in wolf-related issues, need for livestock protection, and fear of the wolf. Perceived harmfulness was in turn correlated with direct and indirect experience with the wolf, and level of education.

4. Our acceptance map predicted decreasing acceptance with increasing altitude of residency and proximity to locations of confirmed wolf presence. This resulted in the overall opposition to the wolf for the Alpine region, albeit substantial regional differences.

5. We found little spatial overlap (6% of Switzerland) between areas where the wolf was accepted and areas of suitable habitat. These areas of socio-ecological suitability were concentrated in the Jura Mountains and in the eastern and southern Alps, and were absent in the western and central Alps. Particularly in the Jura region, which is yet to be colonised, management of human acceptance will be a crucial conservation target.

6. *Synthesis and applications.* We developed an integrative, socio-ecological approach that allowed us to accurately reproduce recent wolf recolonisation. We anticipate our framework to be a powerful tool to reliably evaluate overall suitable habitats and predict short to medium-term range expansion for species whose distribution is also dependent on human attitudes. Because our approach is sensitive to both the ecological and human component, it is ideally suited to identify key regions where proactive and targeted socio-ecological management plans are needed.

Key-words: *Canis lupus*, conservation, habitat suitability model, human-wildlife conflict, large carnivores, range expansion, recolonisation, socio-demographical factors, sociological survey

Introduction

Large carnivores are declining worldwide as a consequence of human-induced habitat loss and degradation, a

declining prey base, and direct anthropogenic persecution (Sanderson *et al.* 2002; Treves & Karanth 2003; Ripple *et al.* 2014). Under these scenarios, several models have been developed to understand and predict animals' distribution and characterise suitable areas (Boyce & McDonald 1999; Hirzel *et al.* 2002; Schadt *et al.* 2002a; Guisan

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& Thuiller 2005; Boyce *et al.* 2015). These habitat suitability models (HSMs) typically use observed animal locations (e.g. sightings) and ecological characteristics (e.g. vegetation type) at and around these locations to infer preferred and suitable habitats (Engler, Guisan & Rechsteiner 2004; Gu & Swihart 2004; Falcucci *et al.* 2009). However, a major limitation of HSMs is the omission of spatially explicit information on the human dimension. This omission can have severe consequences, particularly for large carnivores that often suffer from a negative reputation and direct persecution because of conflict with humans. Excluding the human dimension from HSMs would mean, for example, that a suitable habitat patch in a wildlife-friendly human landscape has the same occupancy potential as a similar suitable patch in a wildlife-hostile landscape.

The human dimension is particularly important for species that encroach into human-dominated landscapes as a consequence of recolonisation and range expansion, giving rise to complex interactions between environmental and anthropogenic factors. For example, recolonisation by large carnivores is likely to be opposed in landscapes dominated by people whose understanding of nature no longer includes the focal species as an integral component. Such unfamiliarity can lead to fear and intolerance because of inability to relate to the presence of an apparently novel component in an idealised version of nature (Kellert *et al.* 1996; Dickman 2010; Chapron *et al.* 2014). Negative attitudes may translate into lobbying and a hostile political microclimate, obstruction to wildlife management plans, or retaliatory killing (Linnell, Swenson & Anderson 2001; Treves & Bruskotter 2014). Positive public attitudes are generally expected in areas where carnivores are absent or where they have always coexisted alongside with people (Kellert *et al.* 1996; Kaczensky, Blazic & Gossow 2004; Majić & Bath 2010). Because human attitude is a crucial factor shaping animals' distribution and habitat use, and influencing long-term persistence of local populations (Mech 1995; Mech & Boitani 2003; Dickman 2010; Chapron *et al.* 2014), its inclusion in HSMs is key to increase model predictive power. This allows a more reliable assessment of occupancy and evaluation of recolonisation potential, and thus the implementation of sound evidence-based actions.

During the past century, large carnivores have been heavily persecuted and locally extirpated throughout continental Europe (Ripple *et al.* 2014). In Switzerland, wolves (*Canis lupus*) were completely eradicated by the end of the 19th century owing to persecution following expansion of human settlements, loss of livestock, and competition with humans for game species (Breitenmoser 1998). By the beginning of the 20th century, across the entire Alpine range, isolated wolf populations persisted only in Italy and in the Balkans (Ciucci *et al.* 2009). As a result of legal protection, during the last decades, the Italian wolf population increased in size and expanded its range from the Apennines to the Western Alps

(Valière *et al.* 2003; Fabbri *et al.* 2007; Marucco *et al.* 2009). The first 'Italian wolf' reached Switzerland in 1995, after over a century of absence from the Swiss territory (Glenz *et al.* 2001), generating turmoil and strong negative reactions particularly amongst rural communities (Hunziker, Hoffmann & Wild-Eck 2001; Wallner & Hunziker 2001).

Today, Switzerland is characterised by vast areas of favourable wolf habitats, which include areas with low human density, intermediate elevations, high prey richness, and the presence of natural land covers (Glenz *et al.* 2001; Marucco & McIntire 2010; Falcucci *et al.* 2013). Under similar favourable natural conditions, the formation of stable and reproducing packs is typically observed 2–7 years after first recolonisation by single long-distance dispersers (Pouille, Lequette & Dahier 1999; Wagner *et al.* 2012; Fabbri *et al.* 2014) with subsequent population expansion. For example, first wolves were observed in France in 1992 and by 2003 the wolf population in the French Alps was estimated to be >120 individuals (Cubaynes *et al.* 2010). Similarly, wolves were first reported in Germany in 1998 and 16 years later, 25 packs lived in the eastern region of the country (Reinhardt *et al.* 2015). Such rapid expansion and increase in numbers are reported also for North America, and are the result of the species' long-distance dispersal ability and reproductive success (Mech, Fritts & Wagner 1995; Boyd & Pletscher 1999; Vilà *et al.* 2003). Despite the prevailing favourable ecological conditions in Switzerland, however, 20 years after the first observed recolonisation event only a handful of single wolves and one well-established pack live in the country (KORA). The recolonisation process of the Swiss territory appears notably slow compared with similar regions in Central Europe. We speculate that the slow recovery of the wolf in Switzerland is largely owing to a low level of acceptance and legal and illegal persecution that outweigh the favourable ecological conditions. At least 15 wolves died of unnatural causes between 1998 and 2015; three were poached, eight were legally shot by the authorities because of conflict with livestock farming, and four were train/road kills (Table 1). This is a remarkable number given as the small wolf population size.

The aim of this study was to integrate both the human and the ecological dimensions in a unified framework to better assess the suitable wolf habitats in Switzerland and better understand the expansion and recolonisation processes. We first investigated the socio-demographical factors related to human acceptance of the wolf. For this purpose, we sent 10 000 questionnaires across all geographical regions in Switzerland. Based on the responses to our questionnaires, we created a spatially explicit model of human acceptance. Finally, we combined our human acceptance model (HAM) with an HSM to create a final socio-ecological suitability model (SESM) that included both human acceptance and ecological components.

Table 1. Ascertained human-related wolf mortality in Switzerland between 1995 and 2015

Legally shot		Poached		Collisions	
Canton	Year	Canton	Year	Canton	Year
VS*	2000	VS	1998	VS [†]	1999
VS	2000	GR	2014	BE [‡]	2006
GR	2001	GR [§]	2014	TI [‡]	2013
VS [¶]	2002			ZH [‡]	2014
VS [¶]	2003				
VS	2006				
VS	2006				
VS	2009				
LU [¶]	2009				
VS	2009				
VS	2010				
VS	2013				
VS [¶]	2015				
VS [¶]	2015				
UR [¶]	2015				

Data from KORA: “Übersichtsliste aller in der Schweiz genetisch nachgewiesenen Wölfe seit 1998” and PRONATURA: “Geschichte der Rückkehr des Wolfes in die Schweiz”.

VS and BE: western Alps; LU and UR: central Alps; GR: eastern Alps; TI: southern Alps; ZH: Central Plateau.

*Likely deadly injured but body not found.

[†]Snowplough.

[‡]Train.

[§]Mistaken for a fox.

[¶]Granted shooting licence but wolf not found within allowed time window.

Materials and methods

STUDY AREA

This study was conducted across Switzerland, spanning an area of 41 285 km², during 2015. Cultivated areas, pastures, forests, grasslands, bare soil, and rocks on mountain slopes as well as human infrastructure characterise the Swiss landscape. Three distinct topographical regions can be recognised: The Alpine region in the south (max. elevation 4634 m above sea level), the densely populated Central Plateau in the middle, and the Jura Mountains in the north (max. elevation 1720 m above sea level). The Alpine region is in turn divided into western, central, eastern, and southern Alps (Fig. 1). Switzerland’s administration is divided into 26 cantons, which to some extent represent regional, social, and cultural differences.

QUESTIONNAIRE SAMPLING

For targeting our questionnaires, we sampled 10 000 random locations across Switzerland. We assigned to each random location the postcode of the closest settlement (i.e. village, town, or city). Several random locations were assigned to the same settlement; in total, 2580 (63%) settlements were sampled from a total of 4070 Swiss settlements. Those random locations (e.g. on mountain peaks) that were >5 km from the closest settlement were again randomly redistributed within the same topographical region. The choice of the 5 km threshold was based on the radius

of human perception of the spatial environment (Schirpke, Tasser & Tappeiner 2013). This sampling procedure ensured spatially representative sampling of interviewees, as opposed to a density-based sampling method, which would have resulted in oversampling of larger cities.

For each random location/settlement, we randomly selected private households using the official phonebook of Switzerland. To allow for random sampling of age and gender within a household, we requested the adult (≥16 years old) who was born earliest in the annual calendar to fill the questionnaire (Hunziker, Hoffmann & Wild-Eck 2001). We sent questionnaires by mail in the language of the respective region along with a pre-paid envelope for return consignment. Interviewees had the option to answer either by mail or to fill the questionnaire online; both responses were anonymous. A total of 9428 questionnaires were successfully delivered (in some cases, the recipient had moved or died, or the address was incomplete) and we received back 3142 filled questionnaires (33.3%), corresponding to 1757 settlements. Only 5% of the interviewees used the online questionnaire.

QUESTIONNAIRE DESIGN AND CORRELATES OF HUMAN ACCEPTANCE

We scored human acceptance based on six questions (questions A1–A6; for the English version of the questionnaire see Appendix S1, Supporting Information). To evaluate correlates of acceptance, we used models of social acceptance and public attitudes towards carnivores (Wallner & Hunziker 2001; Kaczensky, Blazic & Gossow 2004). Accordingly, we included socio-demographical parameters (e.g. gender, age, education) and eight factors (*knowledge* of the wolf, *interest* in wolf-related issues, *fear* of the wolf, perceived *harmfulness* of the wolf, perceived *spatial proximity* of the wolf, *direct experience* with the wolf, *indirect experience* with the wolf, and need for *livestock protection*) in our *a priori* model (see Appendix S1 for a detailed description of the eight factors). Following Hunziker, Hoffmann & Wild-Eck (2001), Bruskotter, Schmidt & Teel (2007), and Kaczensky, Blazic & Gossow (2004), we formulated two to six questions for each factor (Appendix S1). Responses were either measured on a 5-point Likert scale (Likert 1932) or as multiple-choice answers. We calculated a mean score for each factor assessing the answers to the corresponding questions (Appendix S2).

To assess the internal consistency amongst the answers to the questions representing each factor (Kaczensky, Blazic & Gossow 2004), we calculated the reliability coefficient Cronbach’s alpha (α) (Zeller & Carmines 1980). Alpha values suggested inconsistency ($\alpha < 0.7$) for *direct experience*, *indirect experience*, and *knowledge* (Appendix S2). We therefore used principal component analysis to re-group the questions (G1, G2, G3, H1, H2, D1, D2, D3, and D4) of these three factors as suggested by Hunziker, Hoffmann & Wild-Eck (2001). Five questions (D1, D2, D3, D4, and G1) could not be grouped and were dropped. The other questions (G2, G3, H1, and H2) formed one new group (factor) that we named *experience* with the wolf.

Finally, we performed multiple linear regression analysis to assess the relationship between level of acceptance and the socio-demographical parameters and six remaining factors. Since the factor *harmfulness* was highly correlated with other explanatory variables, we additionally performed linear regression analysis with *harmfulness* as dependent variable. We standardised input variables (Gelman 2008) and selected a single best model from all

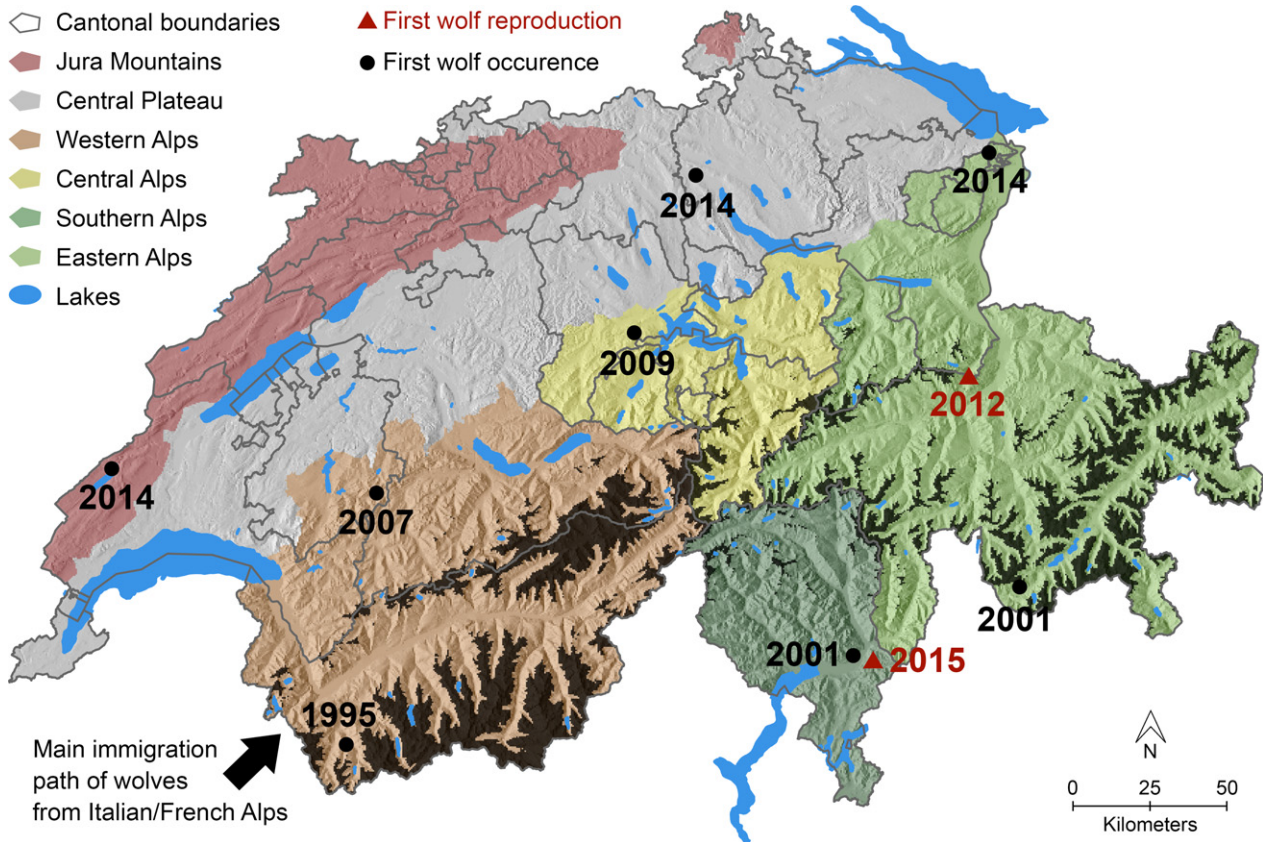


Fig. 1. Topographical map of Switzerland subdivided into Jura Mountains, Central Plateau and Alps. The Alps are in turn divided into four regions. Locations and dates of first confirmed wolf appearance (round symbols) and of first confirmed reproduction (triangles) in different regions are shown. Black-shaded areas denote mountains >2500 m above sea level and may represent a natural expansion corridor from south-west (where wolves enter Switzerland) to north-east (where wolves reproduced in 2012 for the first time).

candidate models based on Akaike's Information Criterion (AIC) (Burnham & Anderson 2002). We performed all statistical and spatial analyses in R (R Core Team 2015).

MODELLING SPATIAL DISTRIBUTION OF HUMAN ACCEPTANCE

We used the first question of the questionnaire 'Are you in favour or against wild living wolves in Switzerland?' to determine human acceptance. Possible answers with assigned levels of acceptance (*LoA*) were as follows: in favour (1), somewhat in favour (0.75), neither nor (0.5), somewhat against (0.25) and against (0). Because some settlements were sampled multiple times (i.e. more than one questionnaire was sent to the same settlement), we calculated a mean level of acceptance for each settlement. We used settlement-specific *LoA* values as the response variable in a logistic regression analysis for proportional data (Guisan, Edwards & Hastie 2002; Zuur *et al.* 2009). We used geo-referenced demographical, social and geographical parameters to create a spatially explicit HAM for the entire study area. After accounting for collinearities, the final set of geo-referenced predictors used in the full model were: *human density*, *mean age*, *language*, *tourism intensity*, *small livestock* (sheep and goat abundance), *wolf distance* (distance to nearest location of confirmed wolf presence), *depreddation* (number of depreddated small livestock), *elevation*, and *agricultural areas* (Appendix S3). We did not use *human-related wolf mortality* as predictor because of the

methodological inconvenience of including very few and spatially distinct data points on a nationwide, large-scale analysis. Each predictor was represented as a raster layer with a resolution of 100 m × 100 m (Fig. 2). For each predictor, we calculated the value of each raster cell using a focal function applied over a circular moving window of 5 km radius. For example, the *human density* value of a given raster cell was the sum of the number of people living within a 5 km radius (Appendix S3). Since the response variable displayed spatial autocorrelation (i.e. settlements close to each other exhibited more similar *human acceptance* values than those further apart), we included an autocorrelation covariate (hereafter: *autocovariate*) as an additional predictor in the final model (Dormann *et al.* 2007). We calculated the autocovariate as distance weighted-sum of neighbouring response values (Bardos, Guillera-Arroita & Wintle 2015). We standardised predictors (Gelman 2008) and followed a backward selection procedure based on AIC (Burnham & Anderson 2002) to simplify the full model, which also included two-way interactions.

We evaluated the predictive performance of our HAM on the basis of the area under the receiver operating characteristic curve (AUC) following 10-fold cross-validation (Elith & Leathwick 2009). AUC measures the overall performance of a model; a model that does not perform better than chance has an AUC of 0.5. We then applied the final model to the entire dataset to predict the probability P_{HAM} associated with each raster cell. The higher the values of P_{HAM} the higher the level of acceptance.

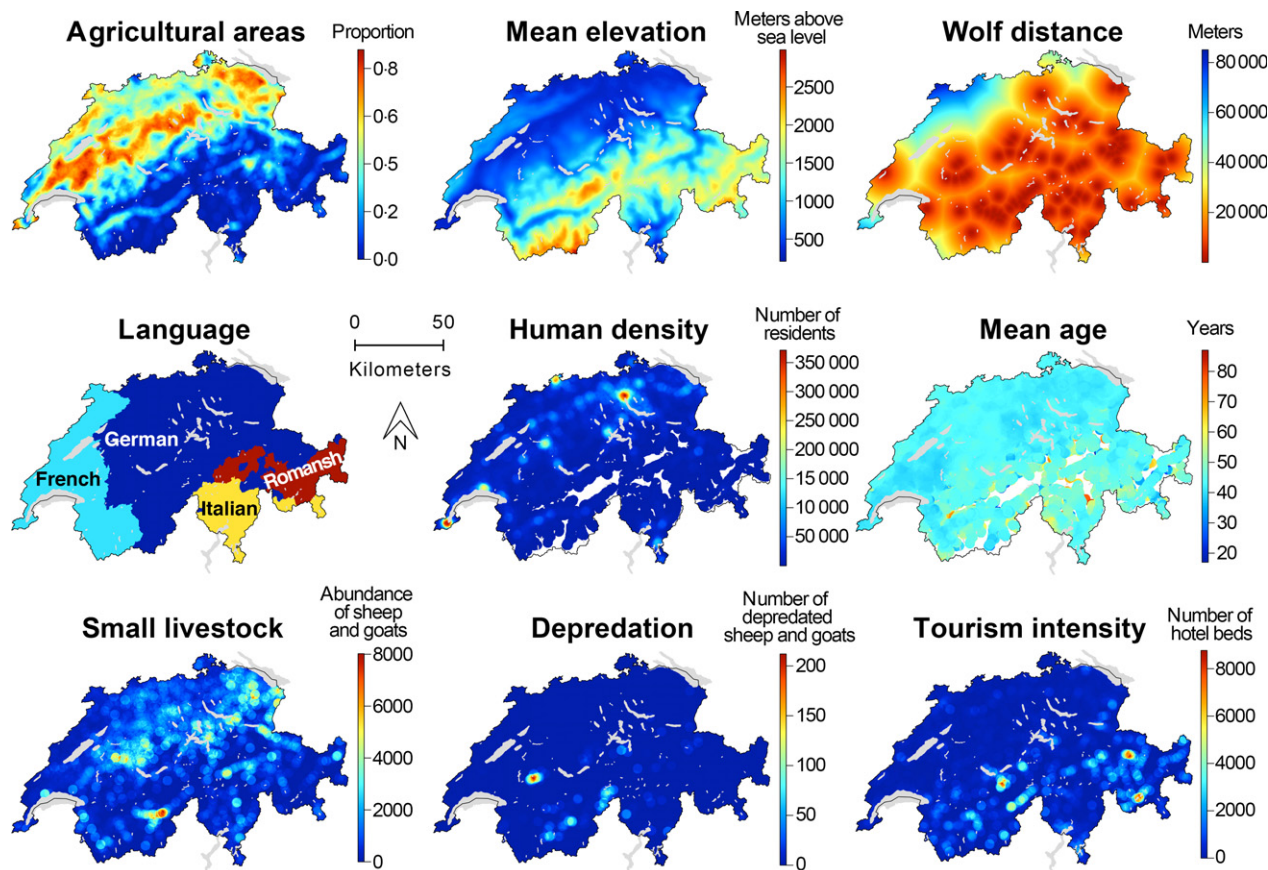


Fig. 2. Raster representation of predictor variables used to model spatial distribution of human acceptance (see text and Appendix S3 for further details).

Raster cells further than 5 km from any settlement (e.g. mountain peaks) were not assigned a P_{HAM} value (Fig. 3a).

CREATION OF A SOCIO-ECOLOGICAL SUITABILITY MODEL

We combined our HAM with a HSM developed by Falcucci *et al.* (2013) to create a final SESM. Falcucci *et al.* (2013) used wolf presence data obtained from the Italian and French Alps, and considered environmental (e.g. land cover), abiotic (e.g. topography) and anthropogenic (e.g. distance to infrastructure) factors to produce a model of the potential distribution of the wolf over the Alpine range. Similar to our HAM, the HSM by Falcucci *et al.* (2013) assigned a continuous suitability value (for clarity renamed P_{HSM}) to each pixel within the entire study area (Fig. 3b). High- P_{HSM} values represented highly suitable areas and low- P_{HSM} values represented unsuitable areas. Because P_{HAM} and P_{HSM} values are not directly comparable in terms of realised suitability (in other words, a pixel with $P_{HAM} = 0$ and $P_{HSM} = 1$ is not necessarily equally as suitable as a pixel with $P_{HAM} = 1$ and $P_{HSM} = 0$), they cannot be multiplied to obtain continuous P_{SESM} values. We therefore transformed the continuous P values of both models to a binary outcome (Fig. 3c and d) – acceptance vs. opposition (for the HAM) and suitable vs. non-suitable (for the HSM) – by setting a model-specific threshold t (Elith & Leathwick 2009). For the HAM, the mean threshold obtained from 10-fold cross-validation at which the sum of the

sensitivity (true positive rate) and specificity (true negative rate) was highest was $t = 0.47$. Predicted P_{HAM} values above this threshold were classified as acceptance of the wolf and values below as opposition to the wolf (Fig. 3c). We obtained threshold values for the HSM from Falcucci *et al.* (2013). We considered intersection between areas where the wolf was accepted ($HAM_{acceptance}$) and areas characterised by a suitable habitat ($HSM_{suitable}$) as overall suitable in our final SESM (Fig. 3e):

$$SESM_{suitable} = HAM_{acceptance} \cap HSM_{suitable}$$

Finally, we evaluated the $SESM_{suitable}$ patches for their potential to sustain wolves based on two criteria (Schadt *et al.* 2002b). Patches were retained and considered overall suitable, either if (i) their area was $\geq 23 \text{ km}^2$, which corresponds to wolves' core areas in central Europe (Ciucci *et al.* 1997; Okarma *et al.* 1998; Kusak, Skrbinšek & Huber 2005; Schmidt, Theuerkauf & Kowalczyk 2007), or if (ii) they were $\leq 20 \text{ km}$ from a patch $\geq 23 \text{ km}^2$. This distance corresponds to the daily average distance moved by wolves in central Europe (Ciucci *et al.* 1997; Jędrzejewski *et al.* 2001; Kusak, Skrbinšek & Huber 2005).

To evaluate the potential effect of future changes in the levels of acceptance on our SESM, we simulated two alternative scenarios. We changed the response of neutral interviewees, from neutral ($LoA = 0.5$) to somewhat in favour ($LoA = 0.75$), and from neutral to somewhat against ($LoA = 0.25$). We then created two alternative SESM (Fig. S1) following the steps outlined above.

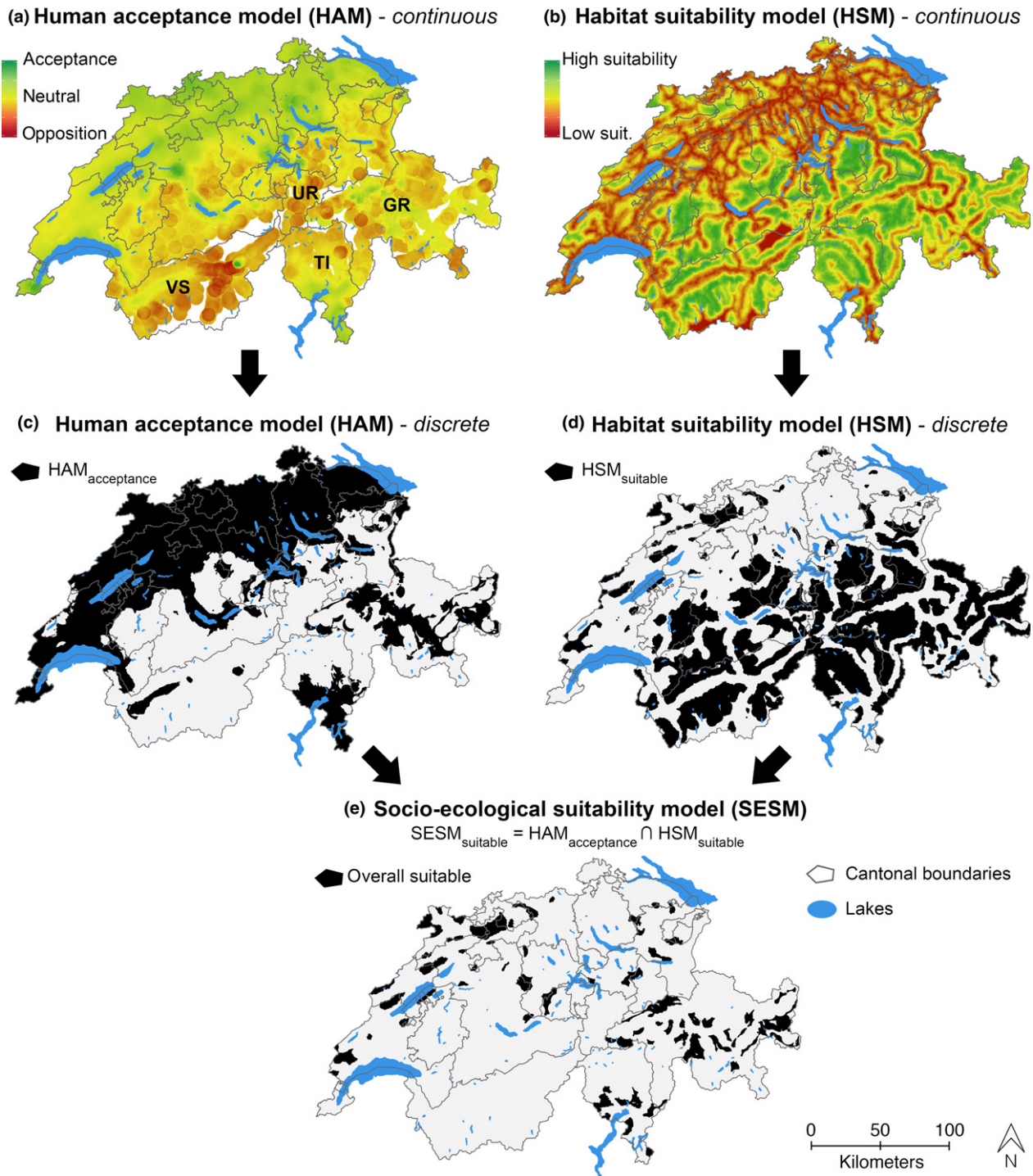


Fig. 3. (a, c) Spatially explicit human acceptance model (HAM), (b, d) habitat suitability model (HSM), and (e) socio-ecological suitability model (SESM) for the wolf in Switzerland (refer to main text for further details). VS, canton of Valais; UR, canton of Uri; TI, canton of Ticino; GR, canton of Graubünden.

Results

CORRELATES OF HUMAN ACCEPTANCE

Our survey revealed an overall polarised but balanced opinion towards the wolf; 49% of the interviewees were against, 45% were in favour and 6% were neutral. After

correcting for human density (note that we intentionally sampled by surface area and not by population density to obtain a representative spatial sampling of interviewees), however, the majority of the Swiss population (59%) was in favour of the wolf whilst only 34% were against it.

Appreciation of the need for livestock protection and interest in wolf-related issues (factors *interest* and *livestock*

protection) were associated with high levels of acceptance of the wolf. On the other side, perception of the wolf as a harmful species and fear towards it (factors *harmfulness* and *fear*) showed a negative relationship with acceptance (Table 2). These four key factors alone explained 78% of the variation in acceptance scores, compared with the 79% of the fully parameterised model that also included perceived *spatial proximity* of the wolf and socio-demographical variables. Older people, men, livestock owners, hunters, and people with only basic education had an overall low level of acceptance; however, the total effect size of these variables was moderate compared with the four above-mentioned key factors (Table 2).

The main driver of *human acceptance* – perceived *harmfulness* of the wolf – was mainly associated with direct and indirect *experience* with the wolf, level of education, membership of an environmental NGO, and hunting activity of a person (Table S1). A negative *experience* and the interviewee being a hunter showed a positive relationship with the perception of the wolf as harmful. Interviewees with a university degree and a pro-environmental orientation, on the other hand, perceived the wolf as beneficial.

SPATIAL DISTRIBUTION OF HUMAN ACCEPTANCE

According to our spatially explicit model, acceptance of the wolf decreased with increasing *mean elevation* and increased with increasing *wolf distance* and *human density* (Table 3). In other words, our model predicted that people living at higher elevations were opposed to the wolf, and people living in densely populated areas or far from

locations of confirmed wolf presence were in favour of the wolf. The interactive effect of *mean elevation* and *small livestock* resulted in low levels of acceptance for the wolf at higher elevation for regions characterised by a high abundance of sheep and goats.

In general, we predicted acceptance of the wolf for the Central Plateau ($\bar{P}_{\text{HAM}} = 0.73$) and Jura Mountains ($\bar{P}_{\text{HAM}} = 0.59$) whilst the Alpine region ($\bar{P}_{\text{HAM}} = 0.42$) showed an overall opposition to the wolf (Fig. 3a and c) albeit with substantial regional differences (Fig. 3a). Predictions of the cantons of Ticino ($\bar{P}_{\text{HAM}} = 0.47$) and Graubünden ($\bar{P}_{\text{HAM}} = 0.44$) in the southern and eastern Alps, respectively, were above average. The cantons of Uri ($\bar{P}_{\text{HAM}} = 0.36$) and Valais ($\bar{P}_{\text{HAM}} = 0.34$) in the central and western Alps, respectively, showed below average values. In terms of predictive performance, our HAM can be regarded as reasonably good (AUC = 0.69).

SOCIO-ECOLOGICAL SUITABILITY MODEL

Our SESM, which considered both human acceptance and suitable habitats, returned a total of 68 patches (totalling 2567 km²; i.e. 6% of Switzerland's area) characterised by the acceptance of the wolf and favourable ecological conditions. 841 km² (33%) were located in the Jura Mountains, 1464 km² (57%) in the Alps, and the remaining 262 km² (10%) in the Central Plateau. This means that only 6% of the entire Alpine region was defined as overall suitable, as opposed to 19% of the Jura Mountains. Despite the favourable ecological conditions (Fig. 3b and d), our SESM did not predict any suitable patches for the western Alps and for most of the central Alps (Fig. 3e).

Table 2. Relation of human acceptance to socio-demographical parameters and factors included in final model (top row) using multiple linear regression analysis. The coefficient estimate and standard error from the final model are reported for each parameter, as is the effect of removing each parameter from the final model on the Akaike's Information Criterion (ΔAIC). The abbreviated explanatory variables are *livest.prot* = need for livestock protection, *hhsz* = size of household, *vac.home* = vacation home owner in mountains, *livest.owner* = livestock owner

Dependent variable: human acceptance				
Model	Estimate	SE	P-value	ΔAIC
proximity + interest + fear + harmfulness + <i>livest.prot</i> + gender + age + children + <i>hhsz</i> + education + hunter + <i>livest.owner</i> + age: <i>livest.owner</i> + children: <i>hhsz</i>				0
– children(yes) (– children(yes): <i>hhsz</i>)	0.27	0.29	0.35	1.1
– <i>hhsz</i> (– children(yes): <i>hhsz</i>)	–0.09	0.22	0.69	1.3
– age: <i>livest.owner</i> (yes)	0.74	0.37	0.05	2.0
– children(yes): <i>hhsz</i>	–1.04	0.47	0.03	3.0
– gender(male)	–0.44	0.19	0.02	3.6
– <i>livest.owner</i> (yes)	–0.43	0.20	0.03	5.4
– education*				8.7
– proximity	–0.58	0.18	<0.001	9.0
– hunter(yes)	–1.21	0.33	<0.001	11.6
– age	–0.83	0.20	<0.001	17.8
– fear	–3.18	0.25	<0.001	155.3
– <i>livest.prot</i>	3.67	0.23	<0.001	235.2
– interest	3.14	0.20	<0.001	238.3
– harmfulness	–6.47	0.28	<0.001	471.7

*Categorical variable with more than two levels.

Table 3. Relation of human acceptance to geo-referenced predictors included in final model (top row) using logistic regression analysis. The explanatory variables are *agricul* = agricultural areas, *elev* = mean elevation, *wolf* = wolf distance, *lang* = language, *dens* = human density, *age* = mean age, *livest* = small livestock, *autocov* = autocovariate

Dependent variable: human acceptance				
Model	Estimate	SE	P-value	Δ AIC
agricul + elev + wolf				0
+ lang + dens + livest				
+ tour + autocov				
+ elev:livest + dens:livest				
+ livest:tour				
– autocov	–0.03	0.12	0.32	1.4
– agricul	–0.27	0.18	0.13	2.2
– lang*				3.8
– dens:livest	0.79	0.43	0.07	4.5
– wolf	0.27	0.12	0.03	4.6
– dens (–dens:livest)	0.39	0.18	0.03	4.7
– tour (–livest:tour)	0.29	0.15	0.06	4.8
– livest:tour	0.76	0.34	0.03	6.3
– elev:livest	–0.93	0.36	0.01	9.9
– livest (–dens:livest)	–0.16	0.13	0.23	25.6
– livest:tour – elev:livest)				
– elev (–elev:livest)	–1.12	0.22	<0.001	39.0

*Categorical variable with more than two levels.

For comparison, the HSM developed by Falcucci *et al.* (2013) predicted 944 km² (22%) of suitable habitat in the Jura Mountains and 12 948 km² (51%) in the Swiss Alps.

Our simulations showed that under the first scenario (i.e. neutral interviewees changed towards somewhat in favour of the wolf), the overall suitable habitat increased by 714 km² (+28%); whilst under the second scenario (i.e. neutral interviewees changed towards somewhat against the wolf) it decreased by 1526 km² (–59%). The difference in the overall suitable area between the two scenarios (2240 km²) corresponded to an area almost as large as the overall suitable area predicted under present conditions (Fig. S1).

Discussion

We integrated human acceptance and ecological components in a unified framework to better assess the suitable wolf habitats in Switzerland. Our results suggested that this integrative approach more precisely represented – at least in the short to medium-term – the wolf recolonisation potential and distribution in Switzerland compared with a model that only considered ecological variables. Specifically, the HSM by Falcucci *et al.* (2013) predicted vast suitable areas evenly distributed across the Swiss Alps (12 948 suitable km²). However, despite 20 years of presence in the territory and high dispersal and recolonisation potential, the vast majority of suitable wolf habitat predicted by HSM has not yet been occupied. It appears that HSM for the wolf in Switzerland has a limited

recolonisation predictive power. Our SESM was drastically more conservative (1464 suitable km²) and better described regional differences. The human dimension captured in this study may indeed be more important than environmental variables for predicting wolf recolonisation and expansion in human-dominated landscapes (Linnell, Swenson & Anderson 2001; Treves & Karanth 2003). This should not come as a surprise given the species' ability to adapt to a variety of different environments (Llaneza, López-Bao & Sazatornil 2012) and the constant persecution it faces (Treves & Karanth 2003). Because our approach is sensitive to both ecological and human component, it represents a powerful tool to identify key regions where proactive and targeted socio-ecological management plans need to be enforced. Including the human component is especially important for sensitive political issues that receive wide social attention, such as the recolonisation of human-dominated landscapes by carnivore species.

The lack of suitable areas predicted by our SESM for the western Alps and for most of the central Alps was mainly driven by a widespread low level of acceptance rather than by the absence of suitable habitats. Interestingly, 20 years after first wolf's appearance, the western Alps have not yet been occupied by the resident wolves. This is even more remarkable considering the vicinity to the well-established wolf population in France and Italy and considering that the western Alps represent the historical and geographical main immigration path to Switzerland (Fig. 1). On the other side, the higher level of acceptance of the wolf in the eastern and southern Alps resulted in vast areas of overall suitable habitat. Here, the presence of pups was confirmed in the cantons of Graubünden (in 2012 and following years) and Ticino (in 2015), 11 and 14 years after first wolf sighting in the respective cantons (KORA).

According to our SESM, the Jura Mountains in the north-western part of the country may represent a second wolf expansion zone, largely because of the overall acceptance and suitable habitat. It is, however, worth noting that this predicted acceptance is largely influenced by distance from locations of confirmed wolf presence (the Jura region has only very recently been in contact with wolves; Fig. 1). Whether or not the present level of acceptance will change as wolves cross the unfavourable highly populated Central Plateau and expand across the Jura Mountains will largely depend on people's attitudes but also on public initiatives to influence acceptance and promote coexistence (Zimmermann, Wabakken & Dötterer 2001; Majić & Bath 2010; Treves, Naughton-Treves & Shelley 2013). In this regard, our integrative approach has the advantage of identifying zones of potential future conflict, thus allowing the development of site-specific management plans to increase the level of human acceptance. The important role of the human dimension on overall habitat suitability is well-illustrated by our simulations, where we showed that even changes in the levels of acceptance in a

small proportion of the population (6% of the interviewees) resulted in a substantial change in overall suitable habitat.

Our HAM confirmed the contrasting attitudes of the rural and urban population to wolf-related issues (Kellert *et al.* 1996). Our results suggest that such contrasting opinions were linked to deep cultural and social differences and to contrasting livestock farming activities across different areas and altitudinal gradient (e.g. traditional pastoralism in remote areas vs. intensive, commercial husbandry activities at lower altitudes). We could not find any systematic information at the desired spatial scale about small livestock husbandry practises and we could therefore not investigate the topic in further detail. As attitudes are dynamic and evolving, and given the underlying social and cultural differences that shape them, proactive and targeted education schemes and awareness campaigns need be tailored to specific conditions and recipients. The effectiveness of such intervention should be regularly evaluated, for example, through repeated surveys to monitor the changes of acceptance over time. Unexpectedly, wolf depredation on livestock was not retained in our final HAM. We believe that this occurred as a result of the nature of the data, which is very much spatially and temporally clumped compared with the large spatial scale of the study (Fig. 2), rather than to a real lack of explanatory power. In fact, despite representing a negligible percentage across all Switzerland, depredation was higher in the western Alps (Table S2).

Low levels of acceptance were mainly driven by perception of the wolf as a harmful species, by fear towards the wolf, and by direct or indirect negative experiences. Similar results were found in a study conducted in the French Alps (Bath 2000). These findings thus highlight three key elements that should be considered if a higher level of acceptance has to be maintained or restored, as in the case of the Jura Mountains and eastern Alps respectively. The belief that the wolf is dangerous to people (47% of interviewees), that it mainly feeds on livestock and not on wild ungulates (33% of interviewees), and that it would not avoid a person upon encounter (22% of interviewees) contributed to the general opposition to the wolf and showed the urgent need to address the 'fear factor' through public education and objective information campaigns. A recent review has highlighted four major avenues of interventions to reduce people's fear of large carnivores and change negative perceptions (Johansson *et al.* 2016). The overall negative connotations linked to the wolf may also largely be associated with media coverage only reporting of livestock depredation events or of alarming scenarios (Treves, Naughton-Treves & Shelley 2013; Fernández-Gil *et al.* 2016). In contrast, information about the positive effects of wolves on prey species or entire ecosystems (e.g. Ripple *et al.* 2001) are only rarely acknowledged and disseminated in Switzerland. According to expectations, perceived spatial proximity to wolves had a negative impact on acceptance (Karlsson & Sjöström 2007), yet this effect was much smaller than the effect size of perceived harmfulness

and fear. In accordance with the other studies, older people were more opposed to the wolf than younger people (Kellert *et al.* 1996; Wallner & Hunziker 2001). This difference may represent genuine generational differences and/or may indicate that young people develop negative attitudes over time. According to expectations, livestock owners, hunters, and people with basic education expressed opposition to the wolf. This confirms how the perception that the wolf has a negative influence on interviewees' personal activities (i.e. harmfulness factor) shapes attitudes (Kellert *et al.* 1996; Ericsson & Heberlein 2003). After correcting for human density, the overall acceptance of the wolf found in our study (59%) was in accordance with the results from a study by Hunziker, Hoffmann & Wild-Eck (2001), who reported 61% positive attitudes for Switzerland. Because younger age classes were under-represented in our sample, our figure appears conservative.

Inferences based on our SESM were centred on the assumption that a low level of acceptance was directly linked to persecution (e.g. shooting), which delays wolf expansion, recolonisation, and settlement. This assumption seemed to hold for the present situation in Switzerland. In the western Alps, where our model predicted the strongest opposition to the wolf, eight wolves were shot (and additional five shooting licences were granted) between 1998 and 2015 (Table 1), mainly because of the predation on livestock. Despite the favourable environmental conditions (Falcucci *et al.* 2013), shooting wolves at a rate of one individual every 2.2 years, may have exacerbated Allee effect (Hurford, Hebblewhite & Lewis 2006) and prevented wolves from settling and reproducing. In contrast, only one wolf was shot in the eastern Alps during the 11 years between first sighting and first reproduction (Table 1). Whilst it appears that shooting prevents (or at least delays) establishment and thus may reduce conflict in the short-term, non-lethal methods have been shown more effective in preventing damage to livestock (Treves, Krofel & McManus 2016) and can lead to considerable changes in attitudes over a short time period (Majić & Bath 2010). On the other hand, besides sending a negative message, legal shooting may encourage poaching as it justifies the use of lethal methods (Chapron & Treves 2016). We advocate that only through implementation of management measures that promote the development of positive attitudes and enhance acceptance, governments can achieve a long lasting coexistence between people and wolves (Carter & Linnell 2016).

In summary, we showed that integrating human attitudes and habitat suitability in a unified SESM reduces predicted suitable area and helps better explain short to medium-term distribution and patterns of recolonisation by the wolf. The remarkable difference in predicted suitable areas returned by the HSM and the SESM was, however, probably partly owing to our integrative approach that discretised both HSM (suitable vs. non-suitable) and HAM (acceptance vs. opposition) instead of using continuous raster values. Future work should identify techniques to precisely weigh HSM and HAM raster values to

allow a more rigorous integrative approach. HSMs are commonly used in conservation practise; hence, their accuracy is crucial for the implementation of sound, evidence-based plans. In light of our findings, we anticipate suitability models for large carnivores to be revisited to include a spatially explicit human component, within the framework presented in this study. We believe that SESMs will become a powerful tool to assess suitable habitats for large carnivores, but also for other species whose distribution is affected by human acceptance, and better understand and predict expansion processes and recolonisation events. Wildlife and humans are increasingly sharing space, and wildlife management planning cannot ignore human attitude and activities. Our approach provides an effective method of accounting for the human and ecological aspects simultaneously, and can be used to develop evidence-based conservation planning, targeting both social and ecological components.

Authors' contributions

G.C., D.M.B. and A.O. conceived the study; G.C. and D.M.B. designed the methodology; D.M.B. collected and analysed the data; D.M.B., G.C. and A.O. interpreted the results; D.M.B. wrote the first version of the manuscript and all authors contributed substantially to the revisions.

Acknowledgements

We thank Marcel Hunziker for his invaluable inputs to the design of our questionnaire, Luigi Maiorano for providing wolf habitat suitability model for the Alpine range, Ronald Schmid, Michael Hässig, and Christian Beglinger for providing geo-referenced data, and the Population Ecology Research Group at UZH for helping in posting 10 000 questionnaires. A special thanks to the KORA office for sharing information on genetically confirmed wolf presence and livestock depredation, and to the 3142 people that returned the completed questionnaire. We thank two anonymous reviewers for valuable comments. This project was funded by University of Zurich research funds to A.O.

Data accessibility

Questionnaire responses are available from the Dryad Digital Repository <https://doi.org/10.5061/dryad.t4t73> (Behr, Ozgul & Cozzi 2017).

Data source of predictors used for a spatially explicit representation of human acceptance is indicated in Appendix S3.

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Received 19 September 2016; accepted 3 February 2017
Handling Editor: Sarah Durant

Supporting Information

Details of electronic Supporting Information are provided below.

Fig. S1. Consequences of change in human acceptance.

Table S1. Relation of perceived harmfulness of the wolf to socio-demographical parameters and factors included in final model (top row) using multiple linear regression analysis.

Table S2. Average abundance of small livestock (sheep and goats) and overall depredation thereof by wolves in the Swiss Alps between 2011 and 2015.

Appendix S1. Questionnaire used in the survey.

Appendix S2. Correlates of human acceptance.

Appendix S3. Predictors used for a spatially explicit representation of human acceptance.